BIODIVERSITAS ISSN: 1412-033X
Volume 25, Number 12, December 2024 **ISSN: 20085-4722** Volume 25, Number 12, December 2024 Pages: 5012-5023 DOI: 10.13057/biodiv/d251239

Mapping of Sapuka Islands water areas, South Sulawesi, Indonesia using Marxan technology for coral reef conservation

JAYA JAYA1,, ROSWIYANTI ROSWIYANTI² , ABDUL RAUF³ , ANDI NUR APUNG MASSISENG⁴ , LUKMAN DARIS⁴ , ST ZAENAB¹

¹Department of Aquaculture, Faculty of Fisheries, Universitas Cokroaminoto Makassar. Jl. Perintis Kemerdekaan Km. 11, Tamalanrea, Makassar 90245, South Sulawesi, Indonesia. Tel.: +62-411-582358, email: yayaremotesensing@ucm-si.ac.id

²Department of Management, Faculty of Social Sciences, Economics and Humanities, Universitas Cokroaminoto Makassar. Jl. Perintis Kemerdekaan Km. 11, Tamalanrea, Makassar 90245, South Sulawesi, Indonesia

³Department for Marine Sciences, Faculty of Fisheries and Marine Science, Universitas Muslim Indonesia. Jl. Urip Sumiharjo No. 225, Makassar 90231, South Sulawesi, Indonesia

⁴Department of Fishery Agrobusiness, Faculty of Fisheries, Universitas Cokroaminoto Makassar. Jl. Perintis Kemerdekaan Km. 11, Tamalanrea, Makassar 90245, South Sulawesi, Indonesia

Manuscript received: 11 September 2024. Revision accepted: 25 December 2025.

Abstract. *Jaya J, Roswiyanti R, Rauf A, Massiseng ANA, Daris L, Zaenab S. 2024. Mapping of Sapuka Islands water areas*, *South Sulawesi, Indonesia using Marxan technology for coral reef conservation. Biodiversitas 25: 5012-5023.* Mapping coastal and marine conservation areas is essential for supporting sustainable resource management and promoting designated utilization practices. This study aimed to assess the condition of coral reefs and develop a map of conservation zones in the Sapuka Islands' waters, South Sulawesi, Indonesia. The study's novelty lies in providing an updated status of coral reef conditions and identifying optimal zones for conservation within the Sapuka Islands. Data were collected through direct observation, Focus Group Discussions (FGD), social mapping, GPS-based coordinate tracking, underwater photo transects, and the identification of planning units, conservation features, and cost metrics. Findings indicated a decline in coral reef area over the past three years, with coverage reducing from 6,094.67 ha in 2022 to 5,771.67 ha in 2023 and further to 5,559.67 ha in 2024. Several coral genera were identified, including *Euphyllia*, *Acropora*, *Montipora*, *Pocillopora*, *Giniastrea*, *Sinularia*, and *Porites*. The Marxan analysis identified a core conservation zone encompassing 6,950 ha, distributed across the waters of Sapuka Caddi Island (6,575 ha) and Pelokang Island (375 ha).

Keywords: Conservation, marine, South Sulawesi, sustainability, zoning

Abbreviations: Marxan: Marine and Spexan

INTRODUCTION

The Sapuka Islands' waters host a rich diversity of natural resources, including various fish species, coral reefs, shellfish, and other marine organisms. However, the abundance and high biodiversity of these coastal and marine resources have led to overexploitation, resulting in environmental degradation (Massiseng et al. 2022). Coastal and marine areas are particularly vulnerable to anthropogenic pressures, which negatively impact resource sustainability (Banikoi et al. 2023; Lyu et al. 2024). Increasing demand for marine and coastal space, driven by the blue economy paradigm, introduces sustainability challenges that necessitate governance transformations and innovative institutional frameworks (Bi et al. 2023). Currently, these areas support aquaculture and capture fisheries activities, primarily utilized by local fishermen (Jaya et al. 2022). Marine spatial planning is essential, serving as a foundation for coastal development, ecosystem protection, and policy formulation to safeguard coastal and marine ecosystems (Liu et al. 2024). The social interaction of fishing communities in the coastal is relatively high in the effort to utilize fishery resources (Daris et al. 2023).

Mapping conservation areas using Marxan technology is a valuable tool for sustainable coastal and marine resource management. This technology facilitates the identification of coral reef conservation zones that can effectively maintain biodiversity and enhance aquatic environmental health. Marxan has become integral in conservation planning, providing users with tools for prioritizing and managing marine conservation areas effectively (Smith 2019; Hanson et al. 2024). Coralreefs, a vital component of the Sapuka Islands marine ecosystem, play a crucial role in sustaining marine biodiversity, supporting human livelihoods, and protecting coastlines (Struebig et al. 2022; Yuan et al. 2024). However, these coral reefs face various threats, including environmental degradation, overfishing, and rising sea surface temperatures, which pose a risk to reef health and survival if not addressed promptly (Mosriula et al. 2018; Obura et al. 2021; Struebig et al. 2022). Additionally, climate stressors and declining water quality have contributed to coral reef deterioration in recent years (Lyons et al. 2020; Andrello et al. 2022). Consequently, activities supporting coral reef conservation are urgently needed.

The South Sulawesi Provincial Government, Indonesia, through its Marine and Fisheries Service, has developed a Coastal and Small Islands Zoning Plan (RZWP3K), which includes designated conservation zones. This plan categorizes waters into several zones: the core zone (habitat protection), the limited-use zone (mutually agreed space), and other zones (including tourism and other uses). Despite these regulatory efforts, technical challenges persist, particularly in the Sapuka Islands, where public awareness and knowledge of conservation practices remain limited. Instances of fishing activities within core zone areas, designated as habitat protection zones, indicate limited compliance with conservation guidelines. To address these challenges, some islands have established local community regulations to govern fishing areas, permissible gear types, and restrictions on fishing activities by non-resident fishers. Sanctions for violations range from fines to social penalties, depending on the severity of the infraction (KKP 2020). The government's conservation zoning efforts could benefit from greater community involvement, as local communities play a central role in compliance and enforcement. Community engagement through participatory mapping can leverage fishermen's local knowledge to identify core zones, utilizing Geographic Information Systems (GIS) and Marxan technology (Kircher et al. 2012; Nunes et al. 2021; Silvano et al. 2023).

This study aims to assess the current state of the coral reef ecosystem and to develop a conservation area map for the Sapuka Islands using Marxan technology. The findings are expected to inform policymakers in the development of conservation strategies for coral reefs, ultimately contributing to sustainable fisheries management.

MATERIALS AND METHODS

Research area and period

The research was conducted in the waters of the Sapuka Islands, located within the Pangkajene and Kepulauan District of South Sulawesi Province, Indonesia, from February to July 2024 (Figure 1). The Sapuka Islands waters, located within South Sulawesi Province, encompass a 50,000-hectare National Marine Conservation Area and a Regional Marine Conservation Area (KKPD) covering 72,493.94 hectares (KKP 2019). The research location within the Pangkajene and Kepulauan Regency includes approximately 66,870 hectares of the Sapuka Islands KKPD. The expansive area designated for marine conservation by the government necessitates alignment with community-based conservation efforts to ensure effective management and foster sustainable coastal and marine resource stewardship.

Data collection

Data types and sources

The primary data for this research includes direct observations, social mapping, coral reef condition surveys, and the recording of GPS coordinate points at data collection sites. Secondary data is sourced from documents provided by relevant agencies, including conservation area data from the South Sulawesi Provincial Maritime and Fisheries Service and coral reef area information from the Makassar Coastal and Marine Resources Management Center (BPSPL). The data types and sources are showed in Table 1.

Figure 1. Map of the Sapuka Islands waters research area, Pangkajene and Kepulauan District, South Sulawesi Province, Indonesia

Table 1. Types and sources of research data

Data types	Data source	Location/unit
Existing condition of coral reefs in Sapuka	Lyzenga GIS analysis, Underwater Photo Transects,	Sapuka Islands waters
Islands waters	CPCE analysis and Social Mapping	
Map of coral reef conservation and rehabilitation	GPS, GIS analysis, Social Mapping, and Marxan	Sapuka Islands waters
areas in Sapuka Islands waters	analysis	

Data collection supports the study's two main objectives. The first objective is to assess the coral reef ecosystem's condition in the Sapuka Islands waters, utilizing data derived from GIS analysis using the Lyzenga method, which visually assesses objects on the sea surface, such as coral reefs and seagrass beds, in addition to underwater photo transects and social mapping. The second objective is to develop a coral reef conservation area map for the Sapuka Islands waters using Marxan technology, informed by field observations, GPS-recorded coordinates, GIS analysis, and social mapping in the study area.

Data collection tools and materials

Photographic documentation was carried out using a digital camera, while coral reef imagery was captured with an underwater Nikon Z6 camera. GPSMAP 65S device was used for recording GPS coordinates of marine conservation sites locations of fishing and aquaculture activities. Outboard motorboats were used for marine transportation to these locations. Banners were prepared for documentation in Focus Group Discussions (FGDs) with local fishermen, along with research location maps for social mapping, notebooks for field data recording, markers for community input on core zones, and scuba diving equipment for underwater research.

Method of collecting data

Before beginning data collection, all research tools and materials were thoroughly checked. GIS and Lyzenga analysis modules were prepared to assess coral reef conditions, along with FGD modules, coral reef documentation protocols for underwater cameras, and data processing modules for Marxan-based conservation area mapping. Field data collection was subsequently undertaken to achieve the study's objectives.

Spatial data collection was conducted to understand coral reef ecosystem distribution, employing GIS analysis using Landsat 8 satellite imagery and the Lyzenga method. Direct observations were performed with scuba diving equipment and underwater cameras, and coordinates were recorded via GPS at conservation area locations, cultivation sites, and fishing sites in the Sapuka Islands. FGDs involving fishermen, community leaders, the Sapuka Islands Village Head, and the Liukang Tangaya Sub-district Head were held alongside social mapping to gather community-driven information on desired conservation zones and areas deemed crucial by local fishers.

Data processing involved integrating coral reef condition data (from Lyzenga analysis and underwater photos), FGD results, and social mapping into ArcGIS. Conservation areas were identified by overlaying official conservation zones with community-defined areas using Marxan technology, providing a comprehensive approach to conservation planning.

Data analysis

Data analysis method for coral reef ecosystem conditions with Lyzenga

Landsat 8 satellite imagery was selected for analyzing coral reefs as it enables the identification of coral reef habitats across both shallow and deep waters. To assess the condition and changes in the extent of coral reef ecosystems in Sapuka Islands waters, Landsat 8 imagery from the last three years (2022, 2023, and 2024) was utilized.

The Lyzenga algorithm, applied via unsupervised classification tools, facilitates analysis of the conditions and spatial changes in coral reef areas. This algorithm specifically calculates the water coefficient ratio (ki/kj), aiding in the assessment of reef habitat conditions. The equation employed for the Lyzenga algorithm, as provided by Philiani et al. (2016), is:

If B5/B2<1 then $log(B2)$ + ki/kj*log (B3) else null

Where \cdot

B5 : Band 5 channel

B2 : Band 2 channel

ki/kj : The coefficient ratio is 1

After generating the resulting image from the Lyzenga algorithm, the next step involves classifying the various marine features, including live coral reefs, dead coral reefs, seagrass distribution, sand distribution, and open sea, using the unsupervised classification system (Irawan et al. 2017). This classification helps to distinguish different habitats and assess the health and distribution of coral reef ecosystems.

Coral reef conservation data analysis method using Marxan technology

Spatial analysis is utilized to examine the distribution of coral reef conservation areas. Geographic phenomena at the research location are assessed using Geographic Information System (GIS) analysis tools (Abd El-Hamid and Hong 2020). GIS is commonly employed to analyze and produce thematic spatial data. For instance, Morandi et al. (2020) identified two conservation areas using GIS in Brazil: The Ecological Corridor (EC) area and the Conservation Unit (CU) in the Cerrado Region.

The development of Marxan technology has advanced by optimizing the algorithm to achieve optimal results at relatively low costs (Moilanen and Kujala 2014). The

Marxan algorithm equation, as described by the Bruce to Milton Transmission Reinforcement Project (2010), is:

$$
\sum Cost + (BLM) \sum Boundary Length + \sum (Species Penalty x SPF) + CTP
$$
\n(a)\n(b)\n(c)\n(d)

Where:

(a) : Planning unit cost is the combined value of socio-economic costs against planning units

(b) : Total Boundary Length is the connectivity value of the planning unit

(c) : Species Penalty is the penalty value given if the biodiversity protection target is not achieved

(d) : Cost Threshold Penalty (TCP) is a penalty solution that incurs costs even if all feature targets are achieved

Based on the results from Marxan technology, the best areas for conservation are identified, as well as areas that provide solutions. Once the selected areas are determined, it is decided whether those waters should be designated for conservation. The areas that provide solutions are the planning units that are selected multiple times, indicating high conservation priority. However, when these areas are selected, the boundaries become less clear, as many selected areas overlap, and decisions must be made regarding their conservation designation.

Planning unit

The study area is divided into smaller parts, known as planning units, each of equal size. Marxan technology works by analyzing the spatial distribution of the planning units. In this study, the area is divided into 1200 planning units, each approximately 1000 hectares. These units have been carefully selected based on various factors, including fishing areas, fish spawning grounds, cultivation locations, and other vital ecosystem data.

Data and scenarios

The development of scenarios in Marxan analysis uses both primary and secondary data. Primary data are obtained through field observations, including coral reef photo data, social mapping data, coordinate points, and other supporting information. Secondary data are gathered from relevant literature to refine and validate the primary data. This data is then converted into spatial form (shapefile) and categorized as polygon type. Proper data preparation is crucial for accurate analysis with Marxan, including the alignment of data resolution and target proportions. If the data have similar spatial resolutions and target proportions, generalization is performed.

Field data collection is another important aspect, where coordinate points are recorded using GPS. These points are analyzed using GIS to produce a spatial dataset relevant to the study's themes. By overlaying multiple maps, the intersection between areas identified by local fishermen and those defined by government regulations can be visualized and analyzed.

Previous studies have successfully applied spatial data analysis to assess marine space usage as an indicator of effective management. For instance, Prestrelo and Vianna (2016) and Bueno and Schiavetti (2019) demonstrated how GIS can help in marine spatial planning. Similarly, participatory GIS approaches have been used in Brazil (Herbst and Hanazaki 2014; Morado et al. 2021; Hettiarachchi et al. 2022) to define conservation area boundaries and improve management effectiveness. GIS also aids in identifying and mapping the condition and distribution of coastal and marine ecosystems, as well as assessing the impact of human activities on coastal areas (Rauf et al. 2020; Jaya et al. 2022).

In Tanzania and Malaysia, participatory geospatial methodologies have been developed for local land-use planning, providing insights into local spatial data and improving community decision-making in formal land-use processes (Zolkafli et al. 2017; Eilola et al. 2019). Additionally, studies by Agaton and Collera (2022) and Schutter et al. (2023) suggest that defining zones for rehabilitation in conservation areas can improve conservation outcomes. By applying these methods, this study aims to provide comprehensive recommendations for the zoning and management of coral reef conservation areas in the Sapuka Islands, ensuring sustainability through informed decision-making processes.

RESULTS AND DISCUSSION

Condition of coral reef ecosystem in Sapuka Islands waters

Condition of coral reef ecosystems in Sapuka Islands waters in 2022

The waters of Sapuka Islands, located within the Coral Triangle, are known for their high biodiversity. Despite this, the coral reef ecosystem in these waters has undergone changes, resulting in a reduction in area. The extent of live coral reefs has decreased, while the area of dead coral reefs has increased. Figure 2 below presents the distribution of coral reef ecosystems in Sapuka Islands waters in 2022.

Condition of coral reef ecosystems in Sapuka Islands waters in 2023

The distribution of coral reefs in Sapuka Islands waters in 2023 showed similar changes as observed in the previous year. The area of live coral reefs decreased, while the extent of dead coral reefs increased. Figure 3 presents the distribution of coral reef ecosystems in Sapuka Islands waters in 2023.

Condition of coral reef ecosystems in Sapuka Islands waters in 2024

The distribution of coral reefs in Sapuka Islands waters in 2024 also underwent changes, similar to those in 2022 and 2023. The area of live coral reefs continued to decrease, while the area of dead coral reefs expanded. Figure 4 illustrates the distribution of coral reef ecosystems in Sapuka Islands waters in 2024.

Figure 2. Distribution of coral reefs in Sapuka Islands waters, South Sulawesi, Indonesia in 2022

Figure 3. Distribution of coral reefs in Sapuka Islands waters, South Sulawesi, Indonesia in 2023

Figure 4. Distribution of coral reefs in Sapuka Islands waters, South Sulawesi, Indonesia in 2024

The GIS and Lyzenga analyses revealed an uneven distribution of coral reefs in the Sapuka Islands waters. Live coral reefs were found to the east and west of Sapuka Besar Island, the west and south of Sapuka Kecil Island, the south of Sambarjaga Island, the west of Pelokang Island, the east of Tinggalungan Island, the east and west of Kembang Lemari Island, the north of Lamuruang Island, and the north and southeast of Cakalangan Island. Dead coral reefs were distributed to the west and south of Sapuka Besar Island, the east and south of Sapuka Kecil Island, the east, west, and south of Sarasang Island, the east, west, and south of Sambarjaga Island, the east and west of Sambargitang Island, the north and west of Pelokang Island, the east and west of Tinggalungan Island, the east and west of Kembang Lemari Island, the north, south, and west of Lamuruang Island, and the east, west, and south of Cakalangan Island.

Several factors contribute to coral reef stress and eventual death, including anthropogenic disturbances, climate change, overfishing, and ocean acidification (De Orte et al. 2021). Additionally, the growth of algae and invertebrates can decrease coral cover, especially in stressed and dying reefs, as these organisms can proliferate on coral tissue (Machendiranathan et al. 2016; Obura et al. 2019; Smith et al. 2021).

Cramer et al. (2022) found that global coral reef ecosystems are most threatened by human-induced climate change. To mitigate this, they recommend that governments implement policies that promote sustainable coral reef

tourism, such as snorkeling and diving, which could support ecosystem conservation.

In addition to coral reefs, seagrass ecosystems were also distributed across various areas: to the north, east, and west of Sapuka Besar Island, west of Sambarjaga Island, and across Pelokang, Tinggalungan, Kembang Lemari, Cakalangan, and Lamuruang Islands. Sand ecosystems were spread across the northern, eastern, western, and southern parts of these islands as well. The GIS and Lyzenga analyses from 2023 indicated similar patterns. Living coral reefs were found to the east and west of Sapuka Besar Island, west, south, and southeast of Sapuka Kecil Island, and across other islands such as Sambarjaga, Sambargitang, Pelokang, and Tinggalungan. Dead coral reefs were similarly found across these locations.

The coral reefs in Sapuka Islands waters exhibit an uneven distribution, with nearly all the islands hosting both living and dead coral reefs. The widespread death of coral reefs globally is attributed to rising sea surface temperatures (Hughes et al. 2018; Foreman et al. 2024), and the degradation is further exacerbated by human activities (Cowburn et al. 2018).

Sims et al. (2024) proposed that measuring the economic value of coral reef conservation, as in the Hawaiian Islands case study, could significantly reduce damage. For example, preserving one meter of coral reef could prevent state losses of \$629 million annually, providing a nature-based solution to coral reef degradation.

In 2024, the GIS and Lyzenga analysis showed the continued presence of living coral reefs in areas similar to those in previous years. These included the east and west of Sapuka Besar Island, the south, north, west, and east of other islands, and dead coral reefs in corresponding locations. The degradation of coral reefs in Sapuka Islands waters is influenced by extreme climate events, such as El Niño, which increases sea surface temperatures, causing heat stress and threatening the health of coral reefs in tropical areas (Romero-Torres et al. 2020).

Study by Sangaji et al. (2024) on Kotania Bay highlighted a decline in coral reef health beginning in 2023. The highest resilience index recorded was 0.74, with the lowest being 0.11. Projections indicate that by 2028, the resilience will decrease to three classes: very low (47.37%), low (42.11%), and moderate (10.53%). By 2032, only two classes will remain: very low (73.68%) and low (26.32%). By 2038, the socio-ecological resilience of coral reefs is expected to be entirely in the very low category (100%).

Sapuka Islands waters hold significant potential for coral reef biodiversity, with genera such as *Euphyllia*, *Acropora*, *Montipora*, *Pocillapora*, *Goniastrea*, *Sinularia*, and *Porites* found throughout the ecosystem. The distribution of these coral reef ecosystems is illustrated in the Figure 5. The condition of coral reefs in Sapuka Islands Waters reveals Pelokang Lompo Island as having the highest coverage of hard coral reefs, with a percentage of 43.33%. Pelokang Caddi Island follows closely at 42.47%, while Cakalangan Island has a coverage of 23.80%. The dominant hard coral species in these waters is *Acropora* (Figure 5.A), characterized by its branched structure, tapered tips, branch stem thickness ranging from 9 mm to 12 mm, and lengths between 55 mm and 81 mm. This type of coral is capable of thriving in both clear and turbid waters at depths of 3 m to 15 m. Another significant coral genus in the area is *Montipora* (Figure 5.B), which is known for its hilly sheet-like formationswith uneven surfaces and white and brown spots. This coral species thrives in clear, flowing waters at depths between 5 m and 15 m.

Coral reef conditions in Sapuka Islands waters, characterized by soft coral cover (Figures 6.A and 6.B), show that Sarassang Island has the highest soft coral coverage, with a percentage of 22.87%. The soft coral species commonly found in these waters include the genus *Goniastrea*, which forms round or elongated colonies with a yellow coloration and sharp surface angles. This species is typically found in coastal areas and relatively shallow waters, as it cannot thrive at depths greater than 50 meters. Additionally, the genus *Pocillopora* was observed, with branching colonies featuring an uneven surface adorned with nodules, thorns, and tentacles. This coral species is capable of growing in both shallow and deeper waters, provided there is adequate sunlight.

Figure 5. Hard coral reef categories of the genus: A. *Acropora*; B. *Montipora*

Figure 6. Soft coral reef categories of the genus: A. *Goniastrea*; B. *Pocillopora*

The condition of the coral reefs in Sapuka Islands Waters, characterized by dead coral reef cover, is shown in Figures 7.A and 7.B. The highest dead coral cover is found on Cakalangan Island, with a percentage of 72.87%. This is followed by Kembang Lemari Island at 59.87%, and Pelokan Lompo Island at 43.53%. The high percentage of dead coral cover on these three islands suggests a significant level of fishing activity using environmentally harmful methods, such as explosives, anesthetics, potassium, and compressors. This continued damage is likely driven by the increasing economic demands of local fishermen.

The condition of the coral reefs in Sapuka Islands Waters, characterized by abiotic covers such as coral fractures (Figure 8.A) and sand (Figure 8.B), shows that Sambarjaga Island has the highest abiotic cover, with a percentage of 67.00%. Lamuruang Island follows with a percentage of 48.73%, and Kembang Lemari Island has 34.33%. The presence of coral fractures is primarily influenced by illegal fishing activities, while the sand cover is largely attributed to sedimentation.

The condition of the coral reefs in Sapuka Islands Waters, with other biota covers such as seagrass beds (Figure 9.A) and starfish (Figure 9.B), shows the highest cover on Sapuka Caddi Island, with a percentage of 15.33%. Both seagrass and starfish ecosystems thrive only in clear waters free from pollution. Therefore, these indicators can be used to assess the level of pollution in the waters.

The results of field observations and CPCe analysis (Table 2 and Figure 10) show that hard coral reefs with moderate conditions are predominantly located on Pelokang Lompo Island (43.33%) and Pelokang Caddi Island (42.47%). Sarassang Island has the highest soft coral cover at 22.87%. Algae cover is most abundant on Sapuka Caddi Island with a percentage of (2.87%) and Pelokang Caddi Island (1.93%). The highest biota cover, including seagrass and starfish, was found on Sapuka Caddi Island (15.33%) and Sarassang Island (6.47%). The presence of soft corals and algae outgrew the new coral growth at the study sites, leading to the coverage of dead corals with algae and soft corals. The composition of coral reef cover in Sapuka Islands Waters was dominated by dead coral and abiotic factors (Figure 10). The high percentage of dead coral can be attributed to rapid algae growth, which covers the reefs still in the process of growth. Additionally, an increase in sea surface temperature, leading to coral bleaching, contributed to the loss of coral. The highest percentage of dead coral was observed on Cakalangan Island (72.87%), followed by Kembang Lemari Island (59.87%) and Pelokang Lompo Island (43.53%). The abiotic cover, which includes coral fractures and sand, was most prominent on Sambarjaga Island (67.00%), Lamuruang Island (48.73%), and Sapuka Caddi Island (38.60%).

Figure 7. Category of dead coral reefs in Sapuka Islands waters, South Sulawesi, Indonesia

Figure 8. Abiotic categories of coral: A. Reefs; and B. Sand

Figure 9. Other biota categories: A. Seagrass beds; and B. Starfish

Figure 10. Composition of coral reef cover in Sapuka Islands waters, South Sulawesi, Indonesia

Table 2. Coral reef cover in Sapuka Islands waters, South Sulawesi, Indonesia

Observation location	Coral reef cover $(\%)$					
	Live hard coral	Soft coral	Dead coral	Abiotic	Alga	Other biota
Pelokan Lompo Island	43.33	3.00	43.53	8.47	0.53	1.13
Pelokan Caddi Island	42.47	2.87	34.60	16.80	1.93	1.33
Sambarjaga Island	6.80	3.00	20.73	67.00	1.20	1.27
Sapuka Caddi Island	2.73	1.27	21.20	38.60	2.87	15.33
Sarassang Island	20.07	22.87	29.13	21.47	0.00	6.47
Lamuruang Island	14.93	0.33	32.73	48.73	0.53	2.73
Cakalangan Island	23.80	1.47	72.87	0.00	0.07	1.80
Kembang Lemari Island	4.40	0.20	59.87	34.33	0.00	1.20

Source: CPCe Survey Results and Analysis (2024)

Comparison of ecosystem area based on Lyzenga analysis of Sapuka Islands waters for the last 3 years

Over the last three years, the coastal and marine ecosystems of the Sapuka Islands Waters have undergone significant changes. These changes are primarily driven by factors such as climate change, fluctuations in physical, chemical, and biological conditions, as well as human activities affecting the region. Table 3 presents the variations in the coastal and marine ecosystems over time.

The satellite image analysis and Lyzenga analysis of the Sapuka Islands' coastal and marine ecosystems from 2022 to 2024 reveal significant changes in habitat areas. The living coral reef ecosystem shrank from 6,094.67 hectares in 2022, decreasing by 323 hectares to 5,771.67 hectares in 2023, and further reducing by 212 hectares to 5,559.67 hectares in 2024.

Figure 11. Map of coral reef conservation and rehabilitation (core zone) in the Sapuka Islands waters, South Sulawesi, Indonesia

Note: Lyzenga survey results and analysis (2024)

Conversely, the dead coral reef area increased significantly, growing from 532.13 hectares in 2022 to 846.13 hectares in 2023 (an increase of 323 hectares), and to 1,058.13 hectares in 2024 with an additional 212 hectares. The seagrass ecosystem also declined, covering 4,453.07 hectares in 2022, then decreasing by 339.05 hectares to 4,114.02 hectares in 2023, and further reducing by 285.48 hectares to 3,828.54 hectares in 2024. Meanwhile, the sand-covered area grew from 3,772.62 hectares in 2022, increasing by 48.56 hectares to 3,821.18 hectares in 2023, and by an additional 35.9 hectares to 3,857.08 hectares in 2024.

These habitat changes are influenced by various oceanographic factors and intensified by fishing activities, which remain crucial to the livelihoods of local fishermen (Hattam et al. 2020). According to research by James et al. (2023), the degradation of coral reef ecosystems leads to

greater hydrodynamic forces impacting nearby bays and increases the risk of erosion due to the loss of ecosystem services. Additionally, McLeod et al. (2021) highlighted that global climate change contributes to coral reef decline, emphasizing the need for resilience assessments to safeguard these ecosystems as climate impacts worsen.

Conservation and rehabilitation of coral reefs in Sapuka Islands waters

The Marxan technology analysis was employed to identify suitable areas for coral reef conservation and rehabilitation in the Sapuka Islands Waters (Figure 11). This technology evaluates the best water areas for conservation by dividing the study area into 1,200 planning units, each covering 1,000 hectares by analyzing and taking into account factors such as fishing grounds, fisheries cultivation zones, and migration patterns of marine species. The results from the Marxan analysis recommended ideal core conservation zones around Sapuka Caddi Island (6,575 hectares) and Pelokang Island (375 hectares).

The designation of waters around Sapuka Caddi Island and Pelokang Island as core conservation zones has been informed by a structured process and scenario analysis using the conservation features embedded in Marxan technology. This selection aligns with the community's vision for the Sapuka Islands Waters, aiming to establish these areas as fish protection zones. Given the favorable condition of the coral reefs, these locations provide essential habitats for fish and marine biota—offering breeding grounds, spawning sites, feeding areas, and shelters from predators.

The strategic alignment of conservation area management with the aspirations of local fishermen, who are primary stakeholders, is crucial for sustaining coastal and marine ecosystems (Qi et al. 2023; Gudka et al. 2024). By involving local communities in planning and safeguarding these zones, the initiative supports both ecological resilience and the livelihoods dependent on these waters, ensuring that conservation measures effectively target ecosystem health and biodiversity while respecting local needs.

ACKNOWLEDGEMENTS

Authors would like to express my gratitude to the Ministry of Education, Culture, Research, and Technology of the Republic of Indonesia for their essential support and funding for the Domestic Cooperation Basic Research (PKDN) program during the 2024 fiscal year. Authors also extend sincere appreciation to the Chancellor of Universitas Cokroaminoto Makassar, Indonesia, for the unwavering support provided throughout the implementation of this research. Authors thankfully acknowledge the Institute for Research and Community Service (LPPM) at Universitas Cokroaminoto Makassar for their valuable guidance and direction in the technical aspects of the research activities.

REFERENCES

- Abd El-Hamid HT, Hong G. 2020. Hyperspectral remote sensing for extraction of soil salinization in the northern region of Ningxia. Model Earth Syst Environ 6: 2487-2493. DOI: 10.1007/s40808-020- 00829-3.
- Agaton CB, Collera AA. 2022. Now or later? optimal timing of mangrove rehabilitation under climate change uncertainty. For Ecol Manag 503: 119739. DOI: 10.1016/j.foreco.2021.119739.
- Andrello M, Darling ES, Wenger A, Suárez-Castro AF, Gelfand S, Ahmadia GN. 2022. A global map of human pressures on tropical coral reefs. Conserv Lett 15 (1): e12858. DOI: 10.1111/conl.12858.
- Banikoi H, Schlüter A, Manlosa AO. 2023. Understanding transformations in the marine coastal realm: The explanatory potential of theories of institutional change. Mar Policy 155: 105791. DOI: institutional change. Mar Policy 155: 105791. DOI: 10.1016/j.marpol.2023.105791.
- Bi M, Wei G, Zhang Z. 2023. The impact of economics and urbanization on marine fisheries sustainability in Atlantic coastal Africa. Ocean Coast Manag 239: 106596. DOI: 10.1016/j.ocecoaman.2023.106596.
- Bueno PF, Schiavetti A. 2019. The influence of fisherman scale in the resilience of socio-ecological systems: An analysis using Q methodology. Ocean Coast Manag 169: 214-224. DOI: 10.1016/j.ocecoaman.2018.12.008.
- Cowburn B, Samoilys MA, Obura D. 2018. The current status of coral reefs and their vulnerability to climate change and multiple human stresses in the Comoros Archipelago, Western Indian Ocean. Mar Pollut Bull 133: 956-969. DOI: 10.1016/j.marpolbul.2018.04.065.
- Cramer KL, Bernard ML, Bernat I, Gutierrez L, Murphy EL, Sangolqui P, Surrey KC, Gerber LR. 2022. The present and future status of ecosystem services for coral reefs. In: DellaSala D, Goldstein MI (eds). Imperiled: The Encyclopedia of Conservation: Volume 1-3. Elsevier, US. DOI: 10.1016/B978-0-12-821139-7.00177-X.
- Daris L, Massiseng ANA, Fachry ME, Zaenab S, Jaya J, Mustaking M. 2023. Types and forms of fishermen conflicts in the utilization of coastal resources in Maros Regency, South Sulawesi Province. IOP

Conf Ser: Earth Environ Sci 1147: 012019. DOI: 10.1088/1755- 1315/1147/1/012019.

- De Orte MR, Koweek DA, Cyronak T, Takeshita Y, Griffin A, Wolfe K, Szmant A, Whitehead R, Albright R, Caldeira K. 2021. Unexpected role of communities colonizing dead coral substrate in the calcification of coral reefs. Limnol Oceanogr 66 (5): 1793-1803. DOI: 10.1002/LNO.11722.
- Eilola S, Käyhkö N, Ferdinands A, Fagerholm N. 2019. A bird's eye view of my village—developing participatory geospatial methodology for local level land use planning in the Southern Highlands of Tanzania. Landsc Urban Plan 190: 103596. DOI: 10.1016/j.landurbplan.2019.103596.
- Foreman AD, Duprey NN, Yuval M, Dumestre M, Leichliter JN, Rohr MC, Dodwell RCA, Dodwell GAS, Clua EEG, Treibitz T, Martínez-García A. 2024. Severe cold-water bleaching of a deep-water reef underscores future challenges for Mesophotic Coral Ecosystems. Sci Total Environ 951: 175210. DOI: 10.1016/j.scitotenv.2024.175210
- Gudka M, Samoilys M, Musembi P, Aboud SA, Grimsditch G, Mabwa R, Yahya SAS, Osuka KE. 2024. Complex coral reefs offer hope for management in a marine protected area in Zanzibar. Reg Stud Mar Sci 77: 10367. DOI: 10.1016/j.rsma.2024.103667.
- Hanson JO, Schuster R, Morrell N, Strimas-Mackey M, Edwards BPM, Watts ME, Arcese P, Bennett JR, Possingham HP. 2024. Prioritizr: Systematic Conservation Prioritization in R. R package version 8.0.4. https://prioritizr.net/.
- Hattam C, Evans L, Morrissey K, Hooper T, Young K, Khalid F, Bryant M, Thani A, Slade L, Perry C, Turrall S, Williamsoni D, Hughes A. 2020. Building resilience in practice to support coral communities in the Western Indian Ocean. Environ Sci Policy 106: 182-190. DOI: 10.1016/j.envsci.2020.02.006.
- Herbst DF, Hanazaki N. 2014. Local ecological knowledge of Fishers about the life cycle and temporal patterns in the migration of mullet (*Mugil liza*) in Southern Brazil. Neotrop Ichthyol 12 (4): 879-890. DOI: 10.1590/1982-0224-20130156.
- Hettiarachchi CJ, Priyankara P, Morimoto T, Murayama Y. 2022. Participatory GIS-based approach for the demarcation of village boundaries and their utility: A case research of the Eastern Boundary of Wilpattu National Park, Sri Lanka. Intl J Geo-Inf 11 (1): 17. DOI: 10.3390/ijgi11010017.
- Hughes TP, Anderson KD, Connolly SR et al. 2018. Spatial and temporal patterns of mass bleaching of corals in the Anthropocene. Science 359 (6371): 80-83. DOI: 10.1126/science.aan8048.
- Irawan J, Sasmito B, Suprayogi A. 2017. Pemetaan sebaran terumbu karang dengan metode Algoritma Lyzenga secara temporal menggunakan citra Landsat 5, 7, dan 8 (Studi Kasus: Pulau Karimunjawa). J Geodesi Undip 6 (2): 56-61. [Indonesian]
- James RK, Keyzer LM, van de Velde SJ, Herman PMJ, van Katwijk MM, Bouma TJ. 2023. Climate change mitigation by coral reefs and seagrass beds at risk: How global change compromises coastal ecosystem services. Sci Total Environ 857 (Pt 3): 159576. DOI: 10.1016/j.scitotenv.2022.159576.
- Jaya J, Laitte MH, Febri F. 2022. Analysis of suitability and carrying capacity of baronang fish (*Siganus* sp) cultivation in coastal waters of Maros Regency. Akuatikisle: J Akuakultur, Pesisir dan Pulau-Pulau Kecil 6 (1): 51-56. DOI: 10.29239/j.akuatikisle.6.1.51-56. [Indonesian]
- Kementerian Kelautan dan Perikanan (KKP). 2020. Rencana Zonasi Wilayah Pesisir dan Pulau-Pulau Kecil. Pusat Data Statistik dan Informasi Kementerian Kelautan dan Perikanan, Jakarta. [Indonesian]
- Kircher, Matthew WL, Game, Eddie, Segan, Dan. 2012. Introduction to Marxan Course Manual Day 2. Centre for Biodiversity & Conservation Science, University of Queensland, Australia.
- Liu S, Cai F, He Y, Qi H, Rangel-Buitrago N, Liu J, Zheng J. 2024. Integrating marine functional zoning in coastal planning: Lessons from the Xiasha Beach Resort case study. Ocean Coast Manag 249: 107016. DOI: 10.1016/j.ocecoaman.2024.107016.
- Lyons MB, Roelfsema C, Kennedy EV et al. 2020. Mapping the world's coral reefs using a global multiscale earth observation framework. Remote Sens Ecol Conserv 6 (4): 557-568. DOI: 10.1002/rse2.157.
- Lyu Y,Wang W, Zhou Z, Geng Z, Jia H, Lu C, Chen Z, Deng W, Xiong X, Shi R, Li H, Yang Z, Lou Q. 2024. Evaluation of the ecological status of shallow-water coral reefs in China using a novel method and identification of environmental factors for coral decline. Mar Pollut Bull 201: 116227. DOI: 10.1016/j.marpolbul.2024.116227.
- Machendiranathan M, Senthilnathan L, Ranith R, Saravanakumar A, Thangaradjou T, Choudhry SB, Sasamal SK. 2016. Trend in coral-

algal phase shift in the Mandapam group of islands, Gulf of Mannar Marine Biosphere Reserve, India. J Ocean Univ China 15: 1080-1086. DOI: 10.1007/S11802-016-2606-8.

- Massiseng ANA, Awaluddin, Fachry ME, Daris L, Jaya. 2022. Catching season and supply chain fishery resources (*Holothuroidea* sp.) small scale on Sapuka Island Pangkep Regency South Sulawesi. Agrikan 15 (2): 638-649. DOI: 10.29239/j.agrikan.15.2.355-366. [Indonesian]
- McLeod E, Shaver EC, Beger M, Koss J, Grimsditch G. 2021. Using resilience assessments to inform the management and conservation of coral reef ecosystems. J Environ Manag 277: 111384. DOI: 10.1016/j.jenvman.2020.111384.
- Moilanen A, Kujala H. 2014. Zonation—Spatial Conservation Planning Methods and Software. Version 4. User Manual, CBIG Conservation Biology Informatics Group, University of Helsinki, Finland.
- Morado CN, de Andrade-Tubino MF, Araújo FG. 2021. Local ecological knowledge indicates: There is another breeding period in the summer for the mullet *Mugil liza* in a Brazilian tropical bay. Ocean Coast Manag 205: 105569. DOI: 10.1016/j.ocecoaman.2021.105569.
- Morandi DT, de Jesus França LC, Menezes ES, Machado ELM, da Silva MD, Mucida DP. 2020. Delimitation of ecological corridors between conservation units in the Brazilian Cerrado using a GIS and AHP approach. Ecol Indic 115: 106440. DOI: 10.1016/j.ecolind.2020.106440.
- Mosriula, Jaya, Hamsir M. 2018. Inventory of damage to coral reefs ecosystem in waters of Bungkutoko Island, Kendari City and Barrang Lompo Island, Makassar City. Akuatikisle: J Akuakultur, Pesisir dan Pulau-Pulau Kecil 2 (2): 67-75. DOI: 10.29239/j.akuatikisle.2.2.67- 75. [Indonesian]
- Nunes MUS, Cardoso OR, Silvano RAM, Fávaro LF. 2021. Participatory mapping and fishers' knowledge about fish and shrimp migration in a subtropical coastal ecosystem. Estuar Coast Shelf Sci 258: 107412. DOI: 10.1016/j.ecss.2021.107412.
- Obura D, Aeby G, Amornthammarong N et al. 2019. Coral reef monitoring, reef assessment technologies, and ecosystem-based management. Front Mar Sci 6: 580. DOI: 10.3389/fmars.2019.00580.
- Obura D, Gudka M, Samoilys M et al. 2021. Vulnerability to collapse of coral reef ecosystems in the Western Indian Ocean. Nat Sustain 5: 104-113. DOI: 10.1038/s41893-021-00817-0.
- Philiani I, Saputra L, Harvianto L, Muzaki AA. 2016. Pemetaan vegetasi hutan mangrove menggunakan metode Normalized Difference Vegetation Index (NDVI) di Desa Arakan, Minahasa Selatan, Sulawesi Utara. Surya Octagon Interdiscip J Technol 1 (2): 211-222. [Indonesian]
- Prestrelo L, Vianna M. 2016. Identifying multiple-use conflicts prior to marine spatial planning: A case research of a multi-legislative estuary in Brazil. Mar Policy 67: 83-93. DOI: 10.1016/j.marpol.2016.02.001.
- Qi W, Hu Y, Linsheng Z, Hui W. 2023. Optimising the relationship between ecological protection and human development through functional zoning. Biol Conserv 281: 110001. DOI: 10.1016/j.biocon.2023.110001.
- Rauf A, Wamnebo MI, Yusuf K. 2020. Application of satellite remote sensing technology in monitoring sediment distribution in the coastal waters of Pangkep Regency, South Sulawesi, Indonesia. AACL Bioflux 13 (5): 3182-3187.
- Romero-Torres M, Acosta A, Palacio-Castro AM, Treml EA, Zapata FA, Paz-García DA, Porter JW. 2020. Coral reef resilience to thermal stress in the Eastern Tropical Pacific. Glob Chang Biol 26 (7): 3880- 3890. DOI: 10.1111/gcb.15126.
- Sangaji M, Louhenapessy DG, Lewerissa YA, Mutmainnah, Lestari F. 2024. Sustainable management of coral reef based on ecology-social resilience level in Kotania Bay, Indonesia. Egypt J Aquat Res 5 (1): 110-116. DOI: 10.1016/j.ejar.2024.03.002.
- Schutter M, ter Hofstede R, Bloemberg J, Elzinga J, van Koningsveld M, Osinga R. 2023. Enhancing survival of ex-situ reared sexual recruits of *Acropora palmata* for reef rehabilitation. Ecol Eng 191: 106962. DOI: 10.1016/j.ecoleng.2023.106962.
- Silvano RAM, Baird IG, Begossi A, Hallwass G, Huntington HP, Lopes PFM, Parlee B, Berkes F. 2023. Fishers' multidimensional knowledge advances fisheries and aquatic science. Trends Ecol Evol 3 (1): 8-12. DOI: 10.1016/j.tree.2022.10.002.
- Sims J, Bausch D, Hoke A, Lindeman C, Kelly M, Zuzak C. 2024. Mapping the risk reduction benefits of coral reef conservation - Hawaiʻi case study. Nature-Based Solutions 5: 100128. DOI: 10.1016/j.nbsj.2024.100128.
- Smith A, Cook N, Cook K, Brown R, Woodgett R, Veron J, Saylor V. 2021. Field measurements of a massive Porites coral at Goolboodi (Orpheus Island), Great Barrier Reef. Sci Rep 11 (1): 15334. DOI: 10.1038/s41598-021-94818-w.
- Smith RJ. 2019. The CLUZ plugin for QGIS: Designing conservation area systems and other ecological networks. Res Ideas Outcomes 5: e33510. DOI: 10.3897/rio.5.e33510.
- Struebig MJ, Aninta SG, Beger M et al. 2022. Safeguarding imperiled biodiversity and evolutionary processes in the Sapuka Islands Center of endemism. BioScience 72 (11): 1118-1130. DOI: 10.1093/biosci/biac085.
- Survey results and analysis. 2024. Primary survey. CPCe.
- Yuan M-H, Lin K-T, Pan S-Y, Yang C-K. 2024. Exploring coral reef benefits: A systematic SEEA-driven review. Sci Total Environ 950: 175237. DOI: 10.1016/j.scitotenv.2024.175237.
- Zolkafli A, Brown G, Liu Y. 2017. An evaluation of Participatory GIS (PGIS) for land use planning in Malaysia. Electron J Inf Syst Dev Ctries 83 (1): 1-23. DOI: 10.1002/j.1681-4835.2017.tb00610.x.