Distribution of phytobenthos in the presence of an invasive alga 
Caulerpa cylindracea on the Algerian west coast

MALIKA GHELLAI1, MOHAMMED EL AMINE BACHIR BOUIADJRA1, LAHOUIAR DAHLOUNM2*,
AHMED MEGHARBI1, BENABDALLAH BACHIR BOUIADJRA3
1Ahmed Zabana University. Relizane, BP 48000, Algeria
2Department of Agronomy, Abdelhamid Ibn Badis University. Av. Hamadou Hossine, 27000, Mostaganem, Algeria.
Tel.: +213-672515273, Fax.: +213-45416827, *email: dahlounhouri@gmail.com
3Laboratoire des Sciences et Techniques de Production Animale, Abdelhamid Ibn Badis University. Av. Hamadou Hossine, 27000, Mostaganem, Algeria


Abstract. Ghellai M, Bachir-Bouiadjra MEA, Dahloum L,Megharbi A, Bachir-Bouiadjra B. 2021. Distribution of phytobenthos in the presence of an invasive alga Caulerpa cylindracea on the Algerian west coast. Biodiversitas 22: 5644-5653. We have targeted and followed the speed of expansion of Caulerpa cylindracea Sonder 1845 through visiting sixteen sites on the Algerian west coast, out in five provinces which are; Tlemcen, Ain Témouchent, Oran, Mostaganem and Chlef, where it was listed in eight stations; the coastal lines of Oran and Mostaganem, and in a single station in the province of Ain Témouchent. The phytosociological surveys were established, made it possible to see the frequency of each identified taxon, mentioning a high degree of disturbance in the appearance of benthic macrophytes regarding the dominance of the invasive species. The calculation of the Shannon index and equitability identify the ecological state of the sites, varies from average (H’ around 3.29 for sites little and not invaded by Caulerpa) to poor (H’: 1.74 for Salamandre), with low macrophytic biodiversity (0.4≤H’≤0.64). The calculation of the Caulerpa racemosa (ICar) index on sandy to rocky substrates with algae does not appear to be a limiting factor for the spread of the invasive species. The previous analyses were validated statistically using Factorial correspondence analysis (FCA), Ascending hierarchical classification (AHC), Principal component analysis (PCA), justifying the unbalanced state according to the degrees of pollution.

Keywords: Algerian, Caulerpa cylindracea, phytosociological

INTRODUCTION

Marine organisms produce a wide variety of natural products, often unique and essential to their survival and ecological performance (Avila et al. 2018; Blunt et al. 2018; Puglisi and Becerro 2018; Carroll et al. 2019). The diversity of algae varies in size and composition and contains a wide variety of species that are widely distributed in different marine sites depending on their adaptability (Kalasariya et al. 2020). Among the natural marine resources, they represent one of the richest sources of natural antioxidants (Khalid et al. 2018). Indeed, algae are known for their antioxidant properties given their richness in bioactive compounds such as polyphenols, terpenoids, alkaloids, saponins, tannins and steroids (Kumar et al. 2019; Ouif et al. 2017; Ouif et al. 2019; Ragunathan et al. 2019). The effects of biological invasions on algal biodiversity are often equivocal (Gribben et al. 2018) with positive and negative impacts on indigenous assembly depending on the geographical context and the temporal and spatial scales of observation (Balestri et al. 2018; Tamburello et al. 2015). Among these bio-invasions, we name Caulerpa cylindracea, exhibiting a strongly invasive behavior during its colonization (Manconi et al. 2020), of Australian origin (Ghellai 2021). However, it has received much less attention than its congeneric species despite being considered one of the most threatening invasions of the Mediterranean Sea (Pazzi et al. 2016). In addition, its presence could have a positive effect on meiofaunal abundance wherever sediment deposition rates are low (Rizzo et al. 2020). From another part, the species presents a source of bioactive compounds of nutritional and pharmacological interest (Belkacemi et al. 2020). Caulerpa cylindracea has settled in several sites on the Algerian coast (Bentaallah and Kerfouf 2013). Research on the expansion and control of this alga has spread over the Algerian west coast by monitoring Bachir Bouiadjra, who has reported its presence on several stations; two sites on the Mostaganémnoise coast; Stidia et Salamandre (Bachir Bouiadjra et al. 2010), and the Arzew site in the city of Oran (Chahrour 2013), as well as on the Kristel coast in 2013, then in Boufser in 2015 in Oran (Hussein and Bensahla 2019).

This study aims to study the degree and speed of expansion of the invasive species C. cylindracea along the Algerian west coast, using index methods and the statistical tool by signaling the danger of its expansion as well that by predicting its future proliferation to the challenge of other benthic macrophytes of ecological and economic interest.

MATERIALS AND METHODS

Choice of study sites
We surveyed 16 affected sites in 5 coastal provinces of western Algeria (Figure 1, Table 1).
frequency of a species, expressed as a percentage, is the ratio of the number of squares where it is present over the total number of squares (Boudouresque 1971). It is given by the following formula:

\[
F = \frac{A}{N} \times 100
\]

The results obtained are classified according to the intervals of values to identify the appearance of which taxon in which the following taxon: If 0.5<F<1.9, the species is considered rare, if 5<F<29.5, the species is designated as present, if 30<F<49.5, the species is common, if 50<F<79.5, the species is frequent in this environment, and if 80<F<100 we are within the limit of having very frequent species.

**Shannon index (H')**

One of the indices derived from the information theory, and the most used in the population’s study is given by the expression of Shannon and Weaver (1949):

\[
H' = - \sum \left[ \frac{P_i}{x \log_2 P_i} \right]
\]

Where, Ri: Recovery of species I; Rt: Total recovery.

According to the obtained values of H’, we can classify the ecological state of the stations according to the following intervals of the thresholds of H’ designated by Simboura and Zenetos (2002): If 0<H’<1.5: we are in a very polluted site with a bad ecological state; If 1.5<H’<3: the site is heavily polluted with a poor ecological status; If 3<H’<4: the site is moderately polluted, the ecological state is designated as medium; If 4<H’<5: a transition zone with a good ecological status; If H’>5: we are in a reference site with very good ecological status.

**The equity index (E) (fairness index)**

To compare the structure of several populations, we use the Regularity (Frontier and Pichod-Vialle 1991) defined as the ratio between effective diversity of the community and its maximum theoretical diversity.
\[ E = \frac{H'}{\log 2} T \]

Where, \( T \): total number of species recorded.
When this tends towards 1, it provides information on an even distribution of dominances between all the species listed; the specific recoveries are inequitably distributed when \( E \) tends towards 0.

The Caulerpa cylindracea Landscape Index; ICar

The ICar landscape index was defined by Cariou et al. (2013) based on 4 criteria: (1) According to the qualification of the colonized substrate; either of (Va) mud, (Sa) sand, (Ro) rock, (RA) rock with algae, (Po) posidonia, (MP) dead matte of posidonia or (CM) coralligenous-maërl; (2) For each site, the area of the colonized zone is qualified in 4 categories either an absence of \( C. racemosa \) (S0), or a colonized spots less than 1 m² (S1), or a colonized spots greater than 1 m² (S2), or a continuous meadow (S3); (3) For each site, the algal cover is qualified in 5 categories: either an absence of \( C. racemosa \) (C0); either a very irregular distribution of the algae at very low density with a number of fronds easy to count (C1); either a greater density of the algae but with an irregular distribution (C2); either a relatively high and regular density with the crawling part (stolon) in one or even two layers (C3); or a significant and regular density with the crawling part in several layers (stolons overlapping) (C4); (4) For each site, the diver's experience is qualified in 3 categories: either he has never seen the Caulerpa (P1), or he has already observed a certain cover signs of \( C. racemosa \) (P2), or he has already observed all the cover indices in \( C. racemosa \) (P3). Therefore, this results in an 8-character code for each site studied.

The results obtained are statistically validated using R software, according to Correspondence Factor Analysis (CFA), Ascending Hierarchical Classification (AHC), and Principal Component Analysis (PCA).

RESULTS AND DISCUSSION

Caulerpa cylindracea frequencies

Our prospecting carried out on the search for the invasive taxon \( C. cylindracea \) on sixteen stations of the Algerian west coast allowed us first of all to record the presence of this taxon in only eight stations among the others, listing from the west to the east; the site of Sbeaat, Bousfer, Kristel, Arzew, Mers El Hadjadj, Stidia, Salamandre, and Bosquet. The overall survey presented in Figure 2 summarizes and groups the status of each station studied with regard to the presence, absence, and estimated frequency of the invasive taxon \( C. cylindracea \).

From the frequencies of \( C. cylindracea \) established in different sites, we could mention the designation of its expansion according to; Frequent and Common at the level of the most affected sites citing; Salamandre, Stidia, Arzew and Bousfer. It is present for the two moderately invaded sites, which are Mers El Hadjadj and Kristel, as well as rare for the sites weakly affected by the invasion, which are Bosquet and Sbeaat.

Bachir (2012), confirmed the critical situation observed at the two stations such as Salamandre and Stidia due to the colonization of \( C. cylindracea \), as well as a weak propagation at the level of the Bosquet site was reported in 2010. Moreover, it was first reported on the Arzew golf course in 2010, and then in 2013 (Bachir et al. 2010b; Chahrour 2013), as well as on the Kristel coast in 2013, then in Bousfer in 2015 (Hussein and Bensahla 2019).

The Diversity index, Shannon index H’, and Fairness E

The Shannon H’ index gives a direct appreciation of the medium. It confirms what is said before about the expansion of the invasive algae; giving for the most affected places the designation of poor concerning the ecological state justified as disturbed, this is for the following sites (Figure 3); Salamandre with a strong spread of the invasive taxon influencing the quality of the marine environment recording the lowest value (\( H' = 1.74 \)) in order to subsequently have a low value for the fairness index (E: 0.4), then the stations strongly affected by \( C. cylindracea \) including; Stidia, Bousfer, Arzew, Mers El Hadjadj, and Kristel (\( H' = 2.09, 2.16, 2.19, 2.48, \) and 2.6, respectively) and which are reported as heavily polluted environments, recording E values ranging from 0.42 to 0.5.

The prospecting carried out at the Stidia and Salamandre stations made it possible to record low values of H’, varying from 1.32 to 1.44, faithfully followed by a fairness index (E) oscillating between 0.73 to 0.76. However, these stations are heavily affected by direct industrial, domestic and urban discharges, and water without prior treatment, thus generating ever-increasing pollution, inevitably affecting the algal community in its diversity (Bachir 2012).

The Sbeaat and Bosquet sites also recorded slightly elevated H’ values of 2.74 and 2.96, respectively, but still designating sites in a state of disturbance, with E values of 0.57 and 0.6, respectively, Bachir (2012), reported that the Bosquet station, which is relatively spared from the pollution of various origins, presents indices of diversity and fairness slightly higher. However, vigilance must be put in this area hence the need to consider appropriate measures before it is too late.

For the three stations visited in the province of Tlemcen, the two stations of Ain Témouchent (Madrid and Bouzedjar), and the last three stations of Chlef, not present the expansion of the studied invasive taxon, allowing us to consider them as moderately polluted sites, with an ecological status designated as medium, showing values H’ between 3.01 and 3.29 (case of Sidna Youchaa and the plage Anglaise stations as retained examples), revealing visibly higher values of E compared to the previous cases, varying between 0.60 and 0.64. Indeed, the diversity index (H’) shows fairly high values in relatively clean stations, as well as lower values in relatively polluted or unstable stations (Chabane 2019).
Study of the *Caulerpa racemosa* Landscape Index; ICar

The exploration results of identifying the visited areas through the *Caulerpa racemosa* index; ICar is summarized in Table 2. ICar was created and applied by Cariou et al. (2013), for the first time in 2011 and 2012, in the Bay of La Revellata (France), over a wide range of depths ranging from 3 to 42 m. The ICar landscape index could be applied throughout the Mediterranean. This simple methodology allows qualifying the different zones colonized by Chlorobiont. It also makes it possible to monitor the spatio-temporal evolution of invasive algae’s distribution (Cariou et al. 2013). The latter covered neither the silt nor the coralligenous, in which it was present on all the other categories of substrates encountered in the studied area (Cariou et al. 2013).

First, the Oued Abdelkah, Sidna Youchaa, Tafsout, Madrid and Bouzedjar sites, located in the extreme west of Algeria, as well as the rest of the stations located in the east of the study area; Oued Elgsob, Marina and the plage Anglaise, are classified as unaffected by the invasive species *C. cylindracea*. The substrates of its studied sites vary from rocky with algae to Sandy.

The ICar study gives us an appreciation of the areas nature affected by *C. cylindracea*, even the nature of the substrate, varying from Sandy to Rocky with algae, as well as a small-scale surface qualification, giving the impression of having spots of less than 1 m², with a very irregular distribution, for sites with low colonization, and spots greater than 1 m² with relatively large distribution, for other sites. In the Salamandre station, continuous meadows with a high density of these invasive algae are found.

The type of substrate does not appear to be a limiting factor in the expansion of *C. cylindracea*, it colonizes *Posidonia*’s dead matte, detritic coast, rocky and sandy substrates, and even the *Posidonia*’s healthy herbarium (Cariou et al. 2013; Gobert and Richir 2019). This index is thought to be applicable and usable by the greatest number of marine biologists, managers of the coastal environment or recreational divers. This is made possible through the dissemination of a brochure identifying the status of substrate colonization by *Caulerpa* (Gobert and Richir 2019).
Statistical study

We established the statistical tests on the basis of Table 3 summarizing the frequencies of recent species at the different study sites.

Correspondence Factor Analysis (CFA)

In order to highlight certain factors which react to the distribution of benthic macrophytes, regarding the invasion of C. cylindracea, factorial correspondence analysis is carried out on a matrix of 21 species, distributed in sixteen sampling stations (Figures 4 and 5).

The F1-F2 plan shows in its negative side, the tolerant species of polluted water, namely respectively Sargassum vulgare, Cystoseira baccata (-2.5), as well as Caulerpa prolifera (-1.48), and C. cylindracea (-1.21), designated in the pollution gradient, followed by Cystoseira stricta and Codium fragile (-1.78); this axis therefore represents through its negative side, a gradient in the intensity of water agitation (Figure 4).

The factorial plane F1-F3 is explained in its negative side by photophile species, namely particularly Jania rubens (-1.61), Caulerpa prolifera (-1.31). The positive side does not explain any gradient. This axis therefore represents, through its negative side, a gradient of brightness. Overall, figures 4 and 5 explain that the expansion of the invasive C. cylindracea species is not limited by the various factors mentioned above, depending on the gradients of pollution, turbidity, and light (Figure 5).

Table 2. Prospecting data according to ICAR in the prospected sites

<table>
<thead>
<tr>
<th>Site</th>
<th>Geographical location</th>
<th>Depth (m)</th>
<th>Year</th>
<th>ICAR</th>
</tr>
</thead>
<tbody>
<tr>
<td>Oued Abdellah</td>
<td>35°05’38” N, 1°51’37” W</td>
<td>2</td>
<td>2020</td>
<td>RA S0 C0 P1</td>
</tr>
<tr>
<td>Sidi Youcha</td>
<td>35°11’81” N, 1°77’95” W</td>
<td>0.5-1</td>
<td>2020</td>
<td>RA S0 C0 P1</td>
</tr>
<tr>
<td>Tafsout</td>
<td>35°10’35” N, 1°39’18” W</td>
<td>1.5</td>
<td>2020</td>
<td>RA S0 C0 P1</td>
</tr>
<tr>
<td>Plage de Madrid</td>
<td>35°18’08” N, 1°23’01” W</td>
<td>0.5-1</td>
<td>2019</td>
<td>Sa S0 C0 P1</td>
</tr>
<tr>
<td>Sheeat</td>
<td>35°18’26” N, 1°8’32” W</td>
<td>0.5-1</td>
<td>2019</td>
<td>Sa S1 C1 P2</td>
</tr>
<tr>
<td>Bouzedjar</td>
<td>35°34’28” N, 1°10’01” W</td>
<td>0.5-1</td>
<td>2019</td>
<td>Sa S0 C0 P1</td>
</tr>
<tr>
<td>Bousfer</td>
<td>35°43’36” N, 0°51’00” W</td>
<td>1-1.5</td>
<td>2019</td>
<td>RA S2 C3 P2</td>
</tr>
<tr>
<td>Kristel</td>
<td>35°49’34” N, 0°29’00” W</td>
<td>1-1.5</td>
<td>2019</td>
<td>Sa S2 C2 P2</td>
</tr>
<tr>
<td>Arzew</td>
<td>35°51’01” N, 0°19’04” W</td>
<td>1-1.5</td>
<td>2019</td>
<td>RA S2 C3 P2</td>
</tr>
<tr>
<td>Mers El Hadjad</td>
<td>35°47’00” N, 0°10’00” W</td>
<td>1-1.5</td>
<td>2019</td>
<td>Sa S2 C3 P2</td>
</tr>
<tr>
<td>S Tudia</td>
<td>35°50’00” N, 0°00’00”</td>
<td>0.5-1</td>
<td>2019</td>
<td>RA S3 C4 P2</td>
</tr>
<tr>
<td>Salamandre</td>
<td>35°56’00” N, 0°05’00” E</td>
<td>0.5-1</td>
<td>2019</td>
<td>RA S1 C1 P2</td>
</tr>
<tr>
<td>Bosquet</td>
<td>36°06’00” N, 0°20’00” E</td>
<td>1-1.5</td>
<td>2018</td>
<td>Sa S0 C0 P1</td>
</tr>
<tr>
<td>Oued Elgsof</td>
<td>36°30’16” N, 1°15’27” E</td>
<td>0.5-1</td>
<td>2018</td>
<td>Sa S0 C0 P1</td>
</tr>
<tr>
<td>Marina</td>
<td>36°30’19” N, 1°40’00” E</td>
<td>0.5-1</td>
<td>2018</td>
<td>RA S0 C0 P1</td>
</tr>
<tr>
<td>Anglaise</td>
<td>36°31’35” N, 1°19’24” E</td>
<td>0.5-1</td>
<td>2018</td>
<td>Sa S0 C0 P1</td>
</tr>
</tbody>
</table>

Table 3. Frequencies (%) of species listed in the different stations

<table>
<thead>
<tr>
<th>Listed species / station</th>
<th>1</th>
<th>2</th>
<th>3</th>
<th>4</th>
<th>5</th>
<th>6</th>
<th>7</th>
<th>8</th>
<th>9</th>
<th>10</th>
<th>11</th>
<th>12</th>
<th>13</th>
<th>14</th>
<th>15</th>
<th>16</th>
</tr>
</thead>
<tbody>
<tr>
<td>Carradoriella elongata</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Caulerpa cylindracea</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Caulerpa prolifera</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Cladophora laetevirens</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Codium fragile</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Corallina officinalis</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Cystoseira baccata</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Cystoseira compressa</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Cystoseira stricta</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Dictyota dichotoma</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Dictyopteris membranacea</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Ellisodiadina elongata</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Hypnea musciformis</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Jania rubens</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Padina pavonica</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Posidonia oceanica</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Sargassum vulgare</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Ulva compressa</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Ulva intestinalis</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Ulva lactuca</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Ulva rigida</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

Note: 1: Oued Abdellah (OA); 2: Sidi Youcha (SY); 3: Tafsout (TF); 4: Madrid (MD); 5: Sheeat (SB); 6: Bouzedjar (BZ); 7: Bousfer (BS); 8: Kristel (KR); 9: Arzew (AZ); 10: Mers El Hadjad (MH); 11: S Tudia (SD); 12: Salamandre (SM); 13: Bosquet (BQ); 14: Oued Elgsof (OG); 15: Marina (MN); 16: Plage Anglaise (AG)
**Ascending Hierarchical Classification (AHC)**

The ascending hierarchical classification of the different sites (Figure 6), shows a gradient of invasion by *C. cylindracea*, highlighting two groups, G1 describing the sites most affected by the invasion of this invasive species, as well as group G2 presenting the sites least affected by the spread of this aggressive taxon.

The hierarchical classification of the 16 stations (Figure 7) allows us to identify two main groups corresponding to the following species assembling: G1: the invasive species *C. cylindracea*; G2: accompanying species (listed).

The second group is subdivided into two subgroups G2” and G2’”, the first (G2”) contains species, which are represented in 1 to 4 sites. The second subgroup (G2’”) adequately explains the taxa that occur at 4 to 15 sites.

The two species located in the extremities are subjected to principal component analysis such as *C. cylindracea* and *P. oceanica*.

**Principal Component Analysis (PCA)**

To confirm our results, a PCA Principal Component Analysis is performed on both ends of the AHC in
particular *C. cylindracea* and *Posidonia oceanica* (Figures 8 and 9). On the other hand, the results of the PCA relating to *C. cylindracea* (Figure 9) reveal a curve that reflects an ever-increasing and unlimited propagation affecting several observation stations, of which the most affected sites by this great change are Stidia and Salamandré.

Our species presents a strong invasive behavior gaining the western Algerian coast (Ghellai et al. 2020), of which it should be noted that this spread of the invasive taxon tends to monopolize the ecological niche of the endemic species *P. oceanica*, even the degradation of a symbolic and endemic ecosystem of the Mediterranean. Long-term interaction with invasive algae could deteriorate the structure of Posidonia meadows, decreasing their initial resistance to invasion due to potential competitive mechanisms between the two macrophytes (e.g. allelopathic effects) (Bernardeau- Esteller 2019).

The characterization of macroalgae and seagrass assemblages shows poor coverage for all species in the invasion area (Bentaallah et al. 2021). In a study area in the Adriatic Sea (Italy), a greater abundance of *C. cylindracea* on the sand and detrital substrates is recorded (Pierucci et al. 2019). This is confirmed by a recent study by Sghaier et al. (2015), along the Tunisian coast, indicating a very high presence of *C. cylindracea* on sandy substrates, as well as the places of rocks and *P. oceanica* meadows. The effect of this alga on benthic communities was described by Piazziet al. (2016), highlighting some main direct and indirect factors affecting the spread of this taxon, many of which are still poorly understood. For example, the relevance of depth, water movement, herbivores and other invaders in the dispersal dynamics of this pest is still unclear (Pierucci et al. 2019). Whereas, these communities play a significant role as environmental indicators because they have different levels of sensitivity to pollution (Kelly et al. 2017).

The expansion process can be very long, showing that only with long-term follow-up studies coupled with better ecophysiological knowledge of *C. cylindracea*, and through manipulation experiments, it might be possible to better understand the key factors invasion of this species in the Mediterranean Sea (Ivesa et al. 2015; Montefalcone et al. 2015). Further studies of biological interest in particular are needed to assess the spread, rate of invasion and impact of this alga, which remarkably affects different areas, at different depths and substrates (Pierucci et al. 2019).

In any case, the conditions of great uncertainty and biological variability, suggest the need for up-to-date monitoring and mapping programs of coastal ecosystems in order to guarantee the survival of endemic species still living in the Mediterranean, and the protection of their marine biodiversity (Bentaallah et al. 2021), which needs to be monitored at different levels of the affected coastline in order to effectively limit its expansion to the detriment of existing algal diversity (Bachir et al. 2021).

Figure 6. Ascending hierarchical classification of different stations
Figure 7. Ascending hierarchical classification of the various recorded species

Figure 8. Plan F1-F2 of the PCA of the Ordination of *Posidonia oceanica*
In conclusion, this study highlights the presence of *C. cylindracea* by prospecting sixteen selected sites on the Algerian west coast in order to estimate its impact on the distribution of other types of plants, even the marine phanerogams: *Posidonia oceanica* and macrophytes, of which it is noted that the sites strongly affected by the invasive taxon, present a low diversity of accompanying algae and suggest their replacement by *C. cylindracea* in particular for the station of Salamandre, Stidia, Arzew, Bousfer, Mers El Hadjadj, kristel, as well as for the Sbeaet and Bosquet sites, where the invasive species is beginning to settle. On the other hand, the stations representing the province of Tlemcen, Ain Témouchent and Chlef, hosting a floral procession rich in macrophytes, where the proportions of frequency, rate of establishment, and dominance of the various identified groups, are in balance.

The degree of danger of this biological pollution, associated with discharges of domestic wastewater, coming from sewerage networks without prior treatment, in certain observation sites, has caused changes and disturbances in these environments, and where the alga *C. cylindracea* has found facilities for colonizing large aquatic surfaces.

Responding to the challenges of mitigating this invasive alien species *C. cylindracea*, we aim; monitor the expansion of this taxon in order to prevent its spread in healthy environments, which is strongly recommended, to carefully monitor these expansions, and to slow them down through awareness-raising operations and the organization of tearing off by professional fishermen, boaters, amateurs, swimmers, to make them know the invasive species, and its effects on the marine ecosystem, in the case where areas occupied by the invasive specimen do not exceed one square meter, therefore the sites weakly affected with tourist potential are the first targets. The monitoring and mapping of the appearance of new colonies must be expressly reported to the services responsible for the protection of the environment and the coast.

**ACKNOWLEDGEMENTS**

We thank warmly Hamsas S, for her help in carrying out this work. We also wish to thank La Direction Générale de la Recherche Scientifique et du Développement Technologique (DGRST, Algeria) for the financial support of the study.

**REFERENCES**


Chahrouh, F. 2013. Etude de la Vitalité des Herbiers à Posidonia oceanica (L), Delile de la Côte Occidentale Algérienne. [Thesis]. Université Ahmed Benbella 1, Oran, Algérie.


