

## Effects of water table level on soil CO<sub>2</sub> respiration in West Kalimantan forested and bare peatland: An experimental stage

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**Abstract.** Astiani D, Burhanuddin, Taherdjadeh M, Curran LM. 2016. *Effects of water table level on soil CO<sub>2</sub> respiration in West Kalimantan forested and bare peatland: An experimental stage*. Nusantara Bioscience 8: 201-206. Peatland forest is one of ecosystems that are susceptible to the alterations of water regimes. Our previous study demonstrated that hydrologic conditions are predominant variables in determining carbon respiration rates from peat soils. However, we have limited empirical measures of how hydrologic conditions (i.e., soil water table) affect the carbon respirations. A study had been conducted to ascertain the effects of water table levels on soil CO<sub>2</sub> respirations of West Kalimantan coastal forested and bare peatland. We simulated natural intact peat core condition by designing 25 of "a double bucket" water level experiment using 5 levels of water tables (0-40cm) and measure soil CO<sub>2</sub> respiration at soil surfaces. Results showed that water table levels significantly affected soil CO<sub>2</sub> respiration, as well as peat microclimate conditions. Results also demonstrated that forest canopy had significant effects on reducing peat CO<sub>2</sub> respiration through its function in maintaining site-environment conditions compared to bare peatland site. This study reveals that it is important to maintain peatland water level close to peat surface and to maintain forest/vegetation covered on peatlands to reduce soil CO<sub>2</sub> emissions.

**Keywords:** Double buckets experiment, landcover changes, microclimate, forested and open peatland

### INTRODUCTION

Peatlands are wetland systems that are susceptible to alterations in their water regimes. It is anticipated that peatlands will be influenced by climate change primarily through modifications in hydrological regimes (Erwin 2009). The prominent impact of global warming on wetland ecosystems is mostly through alterations in precipitation and temperature regimes. Recent trends of climate change are increasing global temperature may also result in increased frequency and intensity of El Niño Southern Oscillation (ENSO) events. Several months of drier conditions may affect several plants and soil microbial processes, including soil emissions of CO<sub>2</sub>, CH<sub>4</sub>, NO, and N<sub>2</sub>O (Sowerby 2008). El Niño events create major water table fluctuations in peatland (Mezhabuddin et al. 2014).

Hydrologic conditions are predominant variables in determining soil respiration rates (Chimner and Ewel 2004). In peatlands, lowering water-table levels have consequence in increasing soil respiration rates in boreal and temperate peatland (Silvola et al. 1996; Chimner and Cooper 2003) because of higher oxygen availability moving into unsaturated peat providing more active transportation and higher aerobic respiration. Current predictions of climate change will alter rainfall patterns in equatorial land Indonesia (Li et al. 2007; IPCC 2013). Alterations in precipitation patterns will have substantial impacts for peatland ecosystems as rainfall and rainfall

patterns are the predominant controls on ecosystem processes such as net primary production (Churkina and Running 1998; Knapp and Smith 2001) and efflux of carbon dioxide from the soil (Davidson and Janssens 2006). In contrast to conditions in upland mineral soils that generally favorable for decomposition, resulting in relatively low carbon densities in peatlands and peatland forests where anaerobic conditions frequently persist, decomposition proceeds much more slowly. However, prolonged drought and land use changes activities could change the pattern of decomposition in these peatlands.

West Kalimantan tropical peatland forests are a seasonal forest type but irregular ENSO events have affected this region for centuries (Curran et al. 1999). Moreover, the development of drainage ditches within the peatland for small to large agriculture and plantation development needs, even non-ENSO years likely affect the hydrological condition of this ecosystem, and thus affect soil carbon emissions. However, unlike boreal and temperate peatlands, we have very limited empirical measures of how hydrologic conditions (e.g., soil water table, water moisture content), as well as soil temperature, influence the soil carbon respirations.

The research objective was to examine the effect of water table on soil carbon respirations in peatland. We also investigated soil microclimate conditions affected by the water table conditions.

## MATERIALS AND METHODS

### Study sites

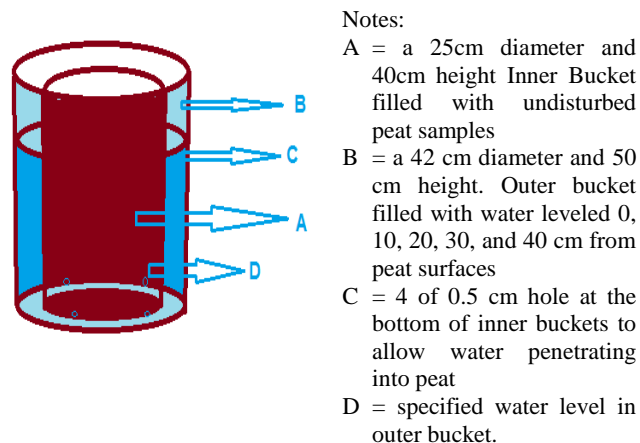
These bucket experiments were conducted in ombrotrophic, coastal type peat in Kubu Raya District, West Kalimantan, Indonesia. The peatland landscape has been influenced by intensive canal development for agriculture purposes. Mean annual rainfall is  $3212 \text{ mm} \pm 529$  (mean  $\pm$  s.d. rainfall data 1990-2014, Supadio Station, Pontianak, West Kalimantan). Average ambient temperature is  $26.5 \pm 0.5^\circ\text{C}$  (mean  $\pm$  s.d. temperature data 1990-2014, Supadio Station), with minimum and maximum temperature  $22.6^\circ\text{C}$  to  $32.2^\circ\text{C}$ . Rainfall data from 1990 has been collected from the closest weather station (Supadio Station) to our site. ENSO (months <100mm rainfall) and non-ENSO year rainfall data are grouped to show the average between both trends and were used as a baseline measure of water level experiment.

### Effect of water table on soil CO<sub>2</sub> emissions

To investigate how soil water table influences the magnitude of peat CO<sub>2</sub> respirations, a set of bucket-experiments were conducted in February - March 2015. We simulated natural peat condition by a series of water table experiment using undisturbed samples of peat under both peat forest environment and opened area. We designed 25 of 'a double buckets' water level experiment using 5 (five) levels of water tables and measure soil CO<sub>2</sub> respiration at soil surfaces on Licor- 8100 soil collars. The double bucket was designed using a 25 cm diameter and 40 cm depth plastic bucket in the inside layer and a 42 cm diameter and 50 cm depth plastic bucket at outside layer. The inside layer bucket was perforated at with 4 of 5mm diameter holes at the bucket base to allow water moving into and draining toward outside bucket to imitate lateral and vertical natural water movement, and filled with 25 cm diameter x 40 cm depth undisturbed peat samples (see Figure 1A and B for illustration). A 25cm-diameter x 40

cm depth, made from stainless steel, peat sampler was designed for taking intact peat samples and let the peat samples undisturbed as in their nature. At the time of peat sample extraction, water levels were between 40-45cm. Based on preliminary data of rainfall and water table measurements taken to set depth levels of these water table treatments, the water depth classes will be divided into 0, 10, 20, 30, and 40 cm.

We applied Complete Random Sampling Method with the five levels of water level depth and 5 replications. The 25 experimental units were considered based on our earlier investigation, the duration in measuring a set of experiment (~2 hours) is optimal to measured them within 2 representative time span in a day (morning and afternoon) to minimize environment condition biases (e.g., peat H<sub>2</sub>O, temperature, time). Complete Random Sampling Design experiments were duplicating for 7-10 days under peatland forest canopy and open area respectively.



**Figure 1.A.** Sketch of regulated water level experiments



**Figure 1.B.** Photos of water level experiment activities

Li-Cor 8100 Automated soil CO<sub>2</sub> flux system (IRGA, Li-Cor 8100, Li-Cor Inc., Lincoln, Nebraska 68504, USA) was used to measured peat CO<sub>2</sub> respiration using 20 cm diameter soil collars. The PVC soil collar is inserted 10 cm beneath soil surface and 2 cm above soil surface and then connected to a Li-Cor 8100-102 soil flux survey chamber. The Li-Cor 8100 contains an infrared gas analyzer to measure soil surface relative humidity, CO<sub>2</sub> and H<sub>2</sub>O concentrations. Attached with auxiliary sensor interface terminal to the Licor 8100 system, we also measured soil temperature with a temperature and water content probes at 10 cm peat depth.

### Data analysis

Throughout the estimation of soil CO<sub>2</sub> respirations during water level experiment, data are presented as mean  $\pm$  SE in selected intervals. One-way ANOVA analysis was used to test the differences of soil CO<sub>2</sub> respirations with water table depth. Soil micro-climate and specific microsite conditions were presented as mean and standard error (SE). Multivariate analyses were used to test the relative importance along with covariance of soil factors that control CO<sub>2</sub> respirations from peat.

## RESULTS AND DISCUSSION

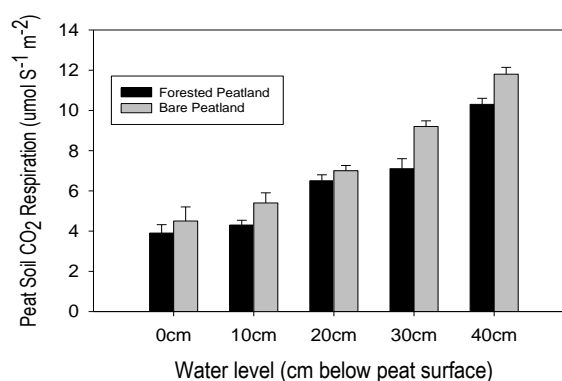
### Preliminary results

The rainfall gaps range widely in months of June through September, 125 and 230 mm consecutively in ENSO and non-ENSO years, with the least amount of 37mm in August in ENSO years, whereas in April and December ENSO year rainfall exceeded non-ENSO years. From our randomly stratified measurements of water table depth in mid-June to September 2013 to 2014, the deepest water table level averaged -30 to -45 cm in forested peatland and in rainy season (November-January) the values ranged from -5 to -30 cm.

Our previous assessment on-site factors variables dictating soil CO<sub>2</sub> along the gradient of land use changes (published separately) indicated that their water table levels influence soil fluxes on each type of land covers such as forest, oil palm, bare, and agricultural on peatlands.

### Water table levels on soil CO<sub>2</sub> respirations

Water levels demonstrated significant effects on peat CO<sub>2</sub> respiration both under forest canopy and opened area (Figure 2). The mean CO<sub>2</sub> flux of water table 0, 10, 20, 30 and 40 cm depth under forest canopy were  $3.9 \pm 0.42$ ,  $4.3 \pm 0.24$ ,  $6.5 \pm 0.3$ ,  $7.1 \pm 0.5$ , and  $10.3 \pm 0.3$   $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$  respectively. Water level 10, 20, 30, and 40 cm increased peat CO<sub>2</sub> respiration of 10%, 67%, 82%, and 164% consecutively when compared to water level of 0 cm. Similar trend occurred on peatland at bare land, which increasing largely when water level was 20cm or further. They were  $4.5 \pm 0.7$ ,  $5.4 \pm 0.5$ ,  $7.0 \pm 0.26$ ,  $9.2 \pm 0.28$ , and  $11.8 \pm 0.34$  respectively for water levels 0, 10, 20, 30, and 40 cm, (see Figure 2). With similar trend to forested peatland, the bare land was increasing CO<sub>2</sub> emission of 20%, 56%, 100% and 162% when water level was increasing from 0 to 40 cm.



**Figure 2.** Peat CO<sub>2</sub> respiration under forest canopy and on bare peatland. One way Anova and Pairwise Multiple Comparison Procedures (Tukey Test) indicated significant difference in soil CO<sub>2</sub> respiration when water level increased to 20 cm and up

When comparing forested peat and bareland, it showed no significant different affecting the peat carbon fluxes ( $t = 0.281$ ;  $df = 8$ ;  $p = 0.786$ ). However, the data trend shows that the respiration in open area was higher than the one under forest canopy (Fig 3). At water level 0 cm on forest peat surface and open peatland, mean peat CO<sub>2</sub> respiration were  $3.9 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$  or  $\approx 54.1 \text{ Mg CO}_2 \text{ ha}^{-1} \text{ y}^{-1}$  and  $4.5 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$  or  $\approx 62.4 \text{ Mg CO}_2 \text{ ha}^{-1} \text{ y}^{-1}$  consecutively. When water level increased to 40 cm, the mean fluxes increased more than double if compared to 0 cm ( $10.3 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$  or  $\approx 143.0 \text{ Mg CO}_2 \text{ ha}^{-1} \text{ y}^{-1}$ ). Moreover, at bare peatland, soil CO<sub>2</sub> fluxes at water level 40 cm reached  $\approx 164 \text{ Mg CO}_2 \text{ ha}^{-1} \text{ y}^{-1}$ .

Our previous study in peatland landscape demonstrate that besides drainage ditches surround the peat landscape, monthly amount of precipitation was also dictated water level in a landscape. In Kubu Raya West Kalimantan, reducing/increasing 100mm rainfall will inflate or deplete water level  $\sim 3.6$  cm (Astiani et al. 2014). Large-scale declines in water table levels during the dry season or on affected drainage peatland deepen the oxic surface peat zone, thereby increasing substrate availability for CO<sub>2</sub> and in turn, releasing decomposition processes. Many biogeochemical processes in the peat are largely governed by the position of the water table (Limpens et al. 2008). The upper peat layer of an undisturbed peatland (range about 5-15 cm) is unsaturated with water and oxic during the dry season and supports most biological activity, whereas the layer below is waterlogged and anoxic. The oxic-anoxic boundary shifts as a result of water table fluctuations, when water table lowered, the oxic layer deepen increasing decomposition processes.

There are many ways in which water can affect ecosystem processes. For example, in very dry conditions, drought can affect the diffusion of enzymes and substrates and basic mobility of micro-organisms. On the other hand, excess water or flooding can create anoxic conditions inhibiting the enzymatic activity allowing only slower and less efficient anaerobic decomposition process to occur (Freeman et al. 2001). Because the buildup of soil organic

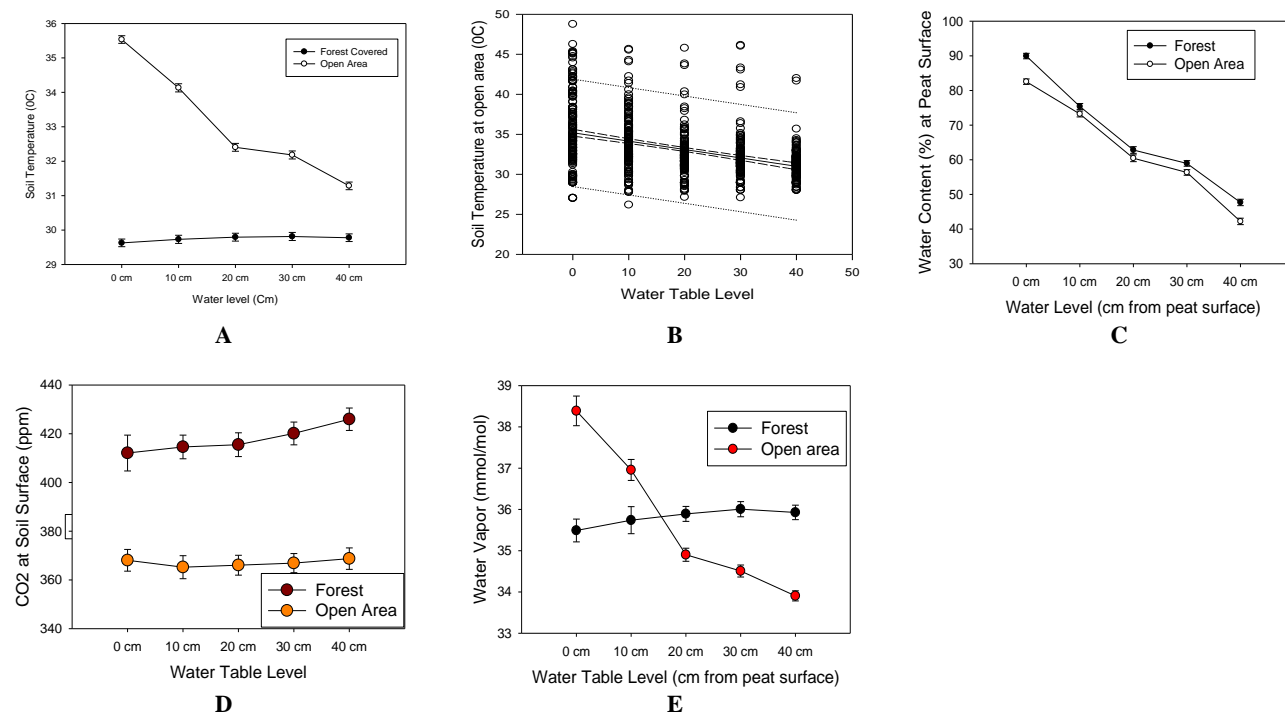
matter in soils is limited by water excess, this process suggests that decomposition is affected more by high water content than by primary production. Therefore on peatland, beside the present of forest cover and their level of degradation (Astiani 2014; 2016), drainage ditches and drought and may be particularly more important to peat hydric systems (Jensen et al. 2003; Davidson and Janssens 2006). Lowered water levels consequently will increase oxygen availability in soil surface and thus will increase biological activities and accelerate rate of organic matter decomposition (Martikainen 1993). The importance of drought in reducing soil carbon stores has also been highlighted in other studies (e.g. Cox et al. 2004; Ciais et al. 2005).

### Effects of water table on soil properties

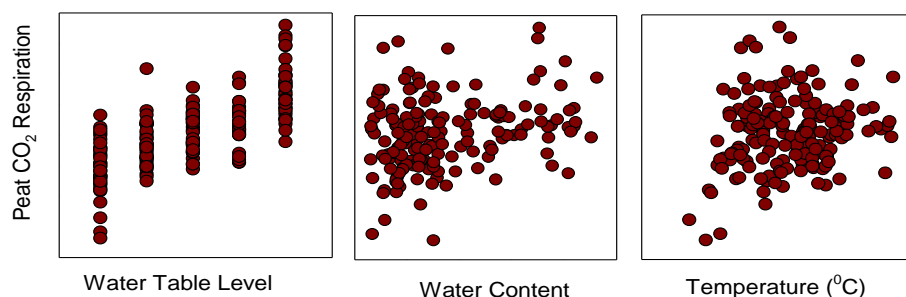
Result from our measurement indicated that soil temperatures on variety water-table level have not fluctuated under forest canopy. On the contrary, soil

temperatures were decreased when water levels were lowered in open area (Figure 3.A). It is indicated that in these water levels range, soil temperature in open area was significantly influence of water table (Figure 3.B) with regression equation: Soil Temperature =  $35.199 - (0.105 * \text{water table level, cm})$ ,  $n = 729$ ,  $r = 0.40$ ,  $R^2 = 0.16$ ,  $p < 0.001$ .

It also indicated that peat soil water content was conducted by water table level. Both forested and bareland sites showed similar trend, yet under forest canopy water content in forest consistently a bit higher than at open one. Soil CO<sub>2</sub> seems similar trend on both sites at variety of water table. However, soil surface CO<sub>2</sub> concentration under forest canopy was higher than open area. It is interesting that water vapor at open area show reducing when water lowered, yet meantime under forest canopy the trend has not occurred. Different behavior on effect of water level on site conditions shows that the combined effect of multi-site factors was mutually influenced each other.



**Figure 3.** A. Soil temperature; B. Soil temperature vs water table level regression; C. Peat water content (%); D. Peat CO<sub>2</sub>; and E. Peat water vapor; on variety of water levels under forest canopy and open area



**Figure 4.** Scatter plot of correlation and multivariate regression of sites factors which significantly influence peat CO<sub>2</sub> emissions

### Effects among site factors on soil respiration

Site factor analysis using Multivariate Linear Regression shows that among site factors, (i.e., water levels, peat surface H<sub>2</sub>O, CO<sub>2</sub>, water content, and temperature), water levels, water content, and temperature were controlling the peat carbon fluxes ( $R^2 = 0.59$ ). The scatter plots and regression are presented in Figure 5. There was strongly positive correlation ( $r = 77\%$ ) among site factors compositely and peat CO<sub>2</sub> respiration. Among-site factors significantly affecting peat CO<sub>2</sub> respiration were water level, water content and soil temperature (Figure 4). Regression analysis demonstrates that these 3 site factors strongly influenced the soil fluxes. The equation model is: (sqrt Peat CO<sub>2</sub> Respiration) =  $1.756 + (0.0297 * \text{Water Table Level}) - (0.0204 * \text{Water Content}) + (0.0220 * \text{Temperature})$ ;  $n = 175$ ;  $r = 0.77$ ;  $R^2 = 0.59$ ; Adj  $r^2 = 0.583$ ; SE = 0.371;  $p < 0.001$ .

### Discussions

There have been many studies in investigating the efflux of carbon from the soil to climate change (e.g. Bellamy et al. 2005; Knorr et al. 2005; Kusler et al. 1999). Many of these studies have primarily focused on responses to temperature (Giardina and Ryan 2000; Davidson and Janssens 2006). However, only a few studies have examined changes in rainfall distributions and patterns on soil carbon efflux (Fay et al. 2000; Jensen et al. 2003; Davidson et al. 2004; Borken et al. 2006). The increased global temperatures result in increased rates of respiration, as highlighted by Davidson and Janssens (2006). However, Schulze and Freibauer (2005) pointed out that decomposition as the effect of soil temperature solely cannot account for trends observed in soil carbon losses. A combination of drought and high temperatures resulted in a reduction in carbon fixation in Europe in 2003, which is equivalent to four years of net carbon storage (Ciais et al. 2005).

Our previous study on peat CO<sub>2</sub> respiration under forest and small scale agriculture on peatland showed that water table consistently act a main significant factors among other (Astiani et al. 2015), even though our research was on smaller spatial scales such as in West Kalimantan peatlands, yet for CO<sub>2</sub> exchange in general, much less variation in exchange rates can be interpreted by environmental and ecological variables (e.g., Bellisario et al. 1999; Moore et al. 1994; Shannon and White 1996). It was also found that changes in the frequency and intensity of weather-related events, such as the local water table lowering during droughts, have a significant consequence on soil CO<sub>2</sub> fluxes. It has been shown that changing water tables give consequence to increase carbon mineralization rates 1.5 to 3 times (Aerts and Ludwig 1997). However, other studies also found small scale variation in relatively high water table levels has shown a minor effect on soil CO<sub>2</sub> flux rates (Chimner and Ewel 2004). Our research, however, is fulfilling a major gap that existed in our knowledge about soil CO<sub>2</sub> flux with variability of peat water table on tropical peatland forests.

Peat water level will control the dynamics of CO<sub>2</sub> through increased soil aeration, which improves soil respiration rates by the higher available oxygen for soil microorganisms (Bouwman 1990). Thus, the impact of soil moisture on soil respiration rates in tropical peatland is more dominant because the seasonal fluctuation in soil temperature in the tropics is relatively low and does not involve low temperature (Inubushi et al. 2003). However, when comparing the respiration rate and site factors among water levels between forested and bare tropical peatland, there were significant magnitudes of responses. Bare peatland consistently emits higher CO<sub>2</sub> than under forested land. These results were supported by Warren et al. (2016) and Hergoualch et al. (2013) that land cover change could impact on increasing soil carbon emission. These results indicate that forest canopy, besides its roles in sequestering atmospheric carbon, can also be beneficial in reducing CO<sub>2</sub> respiration from peatland soil. Therefore, since peatland forest degradation has significant impact on the canopy opening (Astiani 2016), it is important to maintain and enhance the quality of tropical peatland forest at present time.

It is concluded that water level significantly dictating soil CO<sub>2</sub> respiration in peatland landscape. Land cover type change from forests to bare lands was indicated in increasing peatland CO<sub>2</sub> emission. To mitigate the effect on the GHG emission from the half of Indonesian degraded peatland, care should be made to lower the impact on increasing CO<sub>2</sub> in the atmosphere by increasing water level closer to peatland soil surfaces. Peatland hydrology management could help to reduce large amount of CO<sub>2</sub> emitted from this tropical peatland.

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## REFERENCES

- Aerts R, Ludwig F. 1997. Water-table changes and nutritional status affect trace gas emissions from laboratory columns of peatland soils. *Soil Biol Biochem* 29: 1691-1698.
- Astiani D. 2014. Bornean Peatlands: Forest Dynamics, Land Use and Carbon Flux. [Ph.D. Dissertation]. Yale University, USA.
- Astiani D, Mujiman, Hatta M, Hanisa, Fifi F. 2015. Soil CO<sub>2</sub> respiration along annual crops or land-cover type gradients on West Kalimantan degraded peatland forest. *Procedia Env Sci* 28: 132-141.
- Astiani D. 2016. Tropical peatland tree species diversity altered by forest degradation. *Biodiversitas* 17 (1): 102-109.
- Bellamy PH, Loveland PJ, Bradley RI, Lark MR, Kirk GJD. 2005. Carbon losses from all soils across England and Wales 1978-2003. *Nature* 437: 245-248.
- Bellisario LM, Bubier JR, Moore TR. 1999. Control of CH<sub>4</sub> emissions from northern peatland. *Glob. Biogeochem Cycles* 13 (1): 81-91.
- Borken W, Savage K, Davidson EA, Trumbore SE. 2006. Effects of experimental drought on soil respiration and radiocarbon efflux from temperate forest soil. *Glob Chang Biol* 12: 177-193.
- Bouwman AF. 1990. Exchange of greenhouse gases between terrestrial ecosystems and the atmosphere. In: Bouwman (ed) *Soils and the Greenhouse Effect*, John Wiley, New York.
- Churkina G, Running SW. 1998. Contrasting Climatic Controls on the Estimated Productivity of Global Terrestrial Biomes. *Ecosystems* 1: 206-215.
- Chimner RA, Cooper DJ. 2003. Influence of water table position on CO<sub>2</sub> emissions in a Colorado subalpine fen: an in situ microcosm study. *Soil Biol and Biogeochem* 35: 345-351.
- Chimner RA, Ewel KC. 2004. Differences in carbon fluxes between forested and cultivated Micronesian tropical peatlands. *Wetland Ecol Manag* 12 (5): 419-427.
- Ciais PH, Reichstein M, Viovy N. 2005. Europe-wide reduction in primary productivity caused by the heat and drought in 2003. *Nature* 437: 529-533.
- Cox PM, Betts RA, Collins M, Harris PP, Huntingford C, Jones CD. 2004. Amazonian forest dieback under climate-carbon cycle projections for the 21st century. *Theor Appl Climatol* 78: 137-156.
- Davidson EA, Ishida FY, Nepstad. 2004. Effects of experimental drought on soil emissions of carbon dioxide, methane, nitrous oxide, and nitric oxide in moist tropical forest. *Glob Chang Biol* 10: 718-730.
- Davidson EA, Janssens A. 2006. Temperature sensitivity of soil carbon decomposition and feedbacks to climate change. *Nature* 440: 165-173.
- Curran LM, Caniago I, Paoli GD, Astiani D, Kusneti M, Leighton M, Nirarita E, Haeruman H. 1999. Impact of El Niño and logging canopy tree recruitment in Borneo. *Science* 286: 2184-2188.
- Erwin KL. 2009. Wetlands and global climate change: the role of wetland restoration in a changing world. *Wetlands Ecol Manag* 17: 71-84.
- Fay PA, Carlisle JD, Knapp AK, Blair JM, Collins SL. 2000. Altering rainfall timing and quantity in a mesic grassland ecosystem: design and performance of rainfall manipulation shelters. *Ecosystems* 3: 308-319.
- Freeman C, Ostle N, Kang H. 2001. An enzymic 'latch' on a global carbon store. *Nature* 409: 149.
- Giardina CP, Ryan MG. 2000. Evidence that decomposition rates of organic carbon in mineral soil do not vary with temperature. *Nature* 404: 858-861.
- Hergoualch K, Verchot LV. 2013. Greenhouse gas emission factors for land use and land-use change in Southeast Asian peatlands. *Mitig Adapt Strateg Glob Chang* 19:789-807.
- Inubushi K, Furukawa Y, Hadi A, Purnomo E, Tsuruta H. 2003. Seasonal changes of CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O fluxes in relation to land-use change in tropical peatlands located in coastal area of South Kalimantan. *Chemosphere* 52 (3): 603-608.
- IPCC. 2013. Summary of Policymakers. In *Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of Intergovernmental Panel on Climate Change*. In: Stocker TF, Qin D, Plattner GK, Tignor M, Allen SK, Boschung J, Nauels A, Xia Y, Bex V, Midgley PM (eds). Cambridge University Press, Cambridge, UK and New York, NY, USA.
- Jensen K, Beier C, Michelsen A. 2003. Effects of experimental drought on microbial processes in two temperate heathlands at contrasting water conditions. *Appl Soil Ecol* 24: 165-176.
- Knapp AK, Smith MD. 2001. Variation Among Biomes in Temporal Dynamics of Aboveground Prim. *Prod Sci* 291 (5503): 481 - 484.
- Knorr W, Prentice IC, House JI, Holland EA. 2005. Long-term sensitivity of soil carbon turnover to warming. *Nature* 433: 298-301.
- Kusler J, Brinson M, Niering W, Patterson J, Burkett V, Willard D. 1999. Wetlands and climate change: scientific knowledge and management options. White Paper Institute for Wetland Science and Public Policy. Association of Wetland Managers/Wetland International Berne.
- Li WH, Dickinson RE, Fu R, Niu GY, Yang ZL, Canadel JG. 2007. Future precipitation change and their implication for tropical peatlands. *Geophys Res Lett* 34: 1-6.
- Limpens J, Berendse F, Blodau C, Canadell JG, Freeman C, Holden J, Roulet N, Rydin H, and Schaepman-Strub G. 2008. Peatlands and the carbon cycle: from local processes to global implications - a synthesis. *Biogeosciences* 5: 1475-1491.
- Martikainen PJ, Nyklnen H, Grill P, Silvola J. 1993. Effect of a lowered water table on nitrous oxide fluxes on northern peatlands. *Nature* 366: 51-53.
- Mezhabuddin M, Grant RF, Hirano T. 2014. Modelling effects of seasonal variation on water table depth on net ecosystem CO<sub>2</sub> exchange of tropical peatland. *Biogeosciences* 11:577-599.
- Moore TR. 1994. Trace gas emissions from Canadian peatlands and the effect of climatic change. *Wetlands* 14: 223-228.
- Schulze D, Freibauer A. 2005. Carbon unlocked from soils. *Nature* 437: 205-206.
- Shannon RD, White JR. 1996. The effects of spatial and temporal variations in acetate and sulfate on methane cycling in two Michigan peatlands. *Limnology and Oceanography* 41 (3): 435-443.
- Silvola J, Alm J, Ahlholm U, Nykanen H, Martikainen PJ. 1996. CO<sub>2</sub> fluxes from peat in boreal mires under varying temperature and moisture conditions. *J Ecol* 84: 219-228.
- Sowerby A, Emmet BA, Tietema A, Beier, T. 2008. Contrasting effects of repeated summer drought on soil carbon efflux in hydric and mesic heathland soils. *Glob Chang Biol* 14: 2388-2404.
- Warren M, Froking S, Dai Z, Kurnianto S. 2016. Impact of land use, restoration, and climate change on tropical peat carbon stocks in the twenty-first century: Implication for climate mitigation. *Mitig Adapt Strateg Glob Chang*, DOI: 10.1007/s11027-016-9712-1.